

1 **Microbiological and structural aspects of granular sludge from autotrophic**  
2 **denitrifying reactors**

3

4 N. Fernández\*, R. Gómez\*, R. Amils\*, R. Sierra-Alvarez\*\*, J.A. Field\*\* and J.L. Sanz\*

5

6

7 \*Centro de Biología Molecular, Universidad Autónoma, 28049 Madrid, Spain

8 \*\*Department of Chemical and Environmental Engineering, University of Arizona, Tucson, AZ

9 85721-0011, USA

10

11

12 Corresponding author:

13 J. L. Sanz. Dpto. de Biología Molecular, Universidad Autónoma de Madrid, Campus de

14 Cantoblanco, Ctra. de Colmenar, Km.15, Madrid, C.P.: 28049

15 Tel.: +34 91497 8078; fax: +34 91497 8087.

16 E-mail address: [joseluis.sanz@uam.es](mailto:joseluis.sanz@uam.es)

17

18

19

20

21

22

23

24

25

26

27

28

29

30

31 **ABSTRACT**

32 Denitrification is applied in the tertiary treatment of wastewater to reduce N-pollutants.  
33 Fluorescence *in situ* hybridisation (FISH), CARD (catalyzed reporter deposition)-FISH, cloning,  
34 and scanning electron microscopy (SEM) were applied to follow the evolution of the microbial  
35 composition and structure of granular sludge in autotrophic denitrifying bioreactors fed with nitrate  
36 and thiosulfate. With this goal, FISH oligonucleotide probes for the autotrophic denitrifiers,  
37 *Thiobacillus denitrificans* and *Thiomicrospira denitrificans*, were designed and their utility tested.  
38 CARD-FISH and cloning data showed that bacterial diversity changed with bioreactor operation  
39 time. After 110 days of operation, the abundance of *Thiobacillus denitrificans* cells increased  
40 considerably: from 1 to 35% of total DAPI-stained cells and from no isolated clones to 30% of the  
41 total positive clones. This fact strongly suggests that this microorganism played a dominant role  
42 in the autotrophic denitrification. The Archaeal diversity remained almost unchanged and it was  
43 mainly represented by *Methanosaeta soehngenii*. Scanning electron microscopy results indicated  
44 a considerable loss in the integrity of the sludge granules during the operation, with risk of sludge  
45 buoyancy.

46

47

48

49

50

51

52

53

54

55

56

57

58

59

60

61 **KEYWORDS:**

62 Autotrophic denitrification; denitrifying bacteria; granular sludge; molecular ecology;

63 *Thiobacillus denitrificans*; *Thiomicrospira denitrificans*

64

65

66

67

68

69

70

71

72

73

74

75

76

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91 **INTRODUCTION**

92 Denitrification is a microbial redox process by which nitrate is transformed into benign dinitrogen  
93 gas (N<sub>2</sub>). Denitrifying microorganisms can utilize organic or inorganic compounds (i.e. hydrogen  
94 or reduced sulfur compounds) as electron-donors. Heterotrophic denitrifiers are widespread and  
95 phylogenetically diverse. In contrast, few microorganisms appear to be capable of coupling the  
96 reduction of nitrate to N<sub>2</sub> gas to the oxidation of reduced sulfur compounds to sulfate. Among  
97 the obligate autotrophs (i.e. use CO<sub>2</sub> as carbon source) only two species that grow at neutral pH  
98 have been described, *Thiobacillus denitrificans* and *Thiomicrospira denitrificans* (Schedel and  
99 Truper, 1980; Timmer-Ten Hoor, 1981; Kuenen et al., 1992).

100

101 Heterotrophic denitrification (also called nitrate respiration) is widely applied in wastewater  
102 treatment (Green et al., 1995) and other water treatment technologies (Robertson and Anderson,  
103 1999). In contrast, autotrophic denitrification is still underdeveloped for fullscale application.  
104 Significant advantages of autotrophic denitrification, such as its ability to attain simultaneous  
105 removal of nitrate and reduced sulfur compounds (e.g. H<sub>2</sub>S), and the fact that the process does  
106 not require supplementation of organic substrates (e.g. methanol, glucose) as carbon and energy  
107 source, suggest that the process could be attractive for application in the removal of nitrate from  
108 groundwater (Wang and Qun, 2003) and in the treatment of nitrogen and sulfur contamination in  
109 effluents poor in organic matter (e.g. petrochemical wastewaters) (Nugroho et al., 2002; Reyes-  
110 Avila et al., 2004; Sierra-Alvarez et al., 2005), among others.

111

112 Knowledge of the structure of microbial communities in bioreactor biomass is extremely important  
113 to understand any biotechnological process and improve the operation parameters.  
114 Microbiologists have used classical culture-based techniques to achieve this goal. However,  
115 culture-dependent techniques have many shortcomings and, as a consequence, less than 1% of  
116 the microbial diversity has been characterized (Amann et al., 1995). Culturing-independent  
117 molecular biology techniques based on 16S rRNA/rDNA such as fluorescence *in situ* hybridisation  
118 (FISH) and denaturing gradient gel electrophoresis (DGGE) have become essential tools to study  
119 structure and biodiversity of both natural and engineered complex microbial ecosystems (Muyzer  
120 et al., 1996; Amann et al., 2001; Wagner et al., 2003).

121 In the present work, we have applied CARD (catalyzed reporter deposition)-FISH and cloning,  
122 together with scanning electron microscopy (SEM) to follow the evolution of the microbial  
123 composition and structure of granular sludge in laboratory scale autotrophic denitrifying  
124 bioreactors. Given that oligonucleotide probes targeting autotrophic denitrifying microorganisms  
125 are not currently available, specific probes to detect *Thiomicrospira denitrificans* and *Thiobacillus*  
126 *denitrificans* were designed in this study.

127

## 128 **METHODS**

### 129 *Biomass and scanning electron microscopy.*

130 Biomass samples were obtained at different times of operation from two different denitrifying  
131 bioreactors operated under autotrophic conditions. The electron donor for both reactors (R1 and  
132 R2) was thiosulfate ( $S_2O_3^{2-}$ ). R1 and R2 were inoculated with methanogenic granular sludge  
133 obtained from a full-scale upflow anaerobic sludge bed reactor (Eerbeek, The Netherlands). R2  
134 was also supplied with granular activated carbon. Sludge granules were studied by SEM  
135 (Alphenaar et al., 1994). DNA was extracted using FastDNA Spin kit for Soil (BIO101 Inc, Q-  
136 Biogene) according to the manufacturer's protocol.

137

### 138 *Clone libraries.*

139 The 16S rRNA genes from mixed microbial DNA were amplified by PCR. To obtain almost  
140 complete 16S rRNA gene sequences, two oligonucleotide primer pairs were used: 27F and 1492R  
141 (annealing T: 56 °C) for the Bacteria domain and 25F and 1492R (annealing T: 52 °C) for the  
142 Archaea domain (Lane, 1991). The thermal profile for amplification included 10 min at 94 °C for  
143 denaturing, 30 cycles of elongation (94 °C for 1 min, primer annealing at specific temperature for  
144 1 min, and 3 min at 72 °C of elongation), and, finally, 10 min at 72 °C to complete the elongation.  
145 The 16S rRNA gene amplicates (length 1465–1467 bp) were cloned using TOPO Cloning Kit  
146 (Invitrogen, San Diego, CA) and then transformed into competent *E. coli* cells. Subsequently,  
147 representatives of each group were amplified by PCR, using the M13 primer set (Invitrogen), and  
148 sequenced. Sequences were compared with the NCBI databases by using the Basic Local  
149 Alignment Search Tool (BLAST) program to identify the closest sequence.

150

151 *FISH and CARD-FISH.*

152 Oligonucleotide probes design (ARB program, package available at [http://www.mikro.biologie.tu-](http://www.mikro.biologie.tu-muenchen.de/)  
153 [muenchen.de/](http://www.mikro.biologie.tu-muenchen.de/); Fuchs et al., 1998), FISH (Amann et al., 1995, 2001), CARD-FISH (Pernthaler et  
154 al., 2002) and cell counts were performed as previously described. The probes EUB338, ARC915,  
155 NON338, ALF968, BET42a, GAM42a, SRB385, LGC354a, MEB859, MSSH859, MC1109,  
156 MG1200 (described elsewhere) and TBD121 and TMD131 (this work) were tested.

157

## 158 **RESULTS AND DISCUSSION**

### 159 **Oligonucleotide design**

160 Specific oligonucleotide probes for FISH were designed using the ARB package. TBD121 (5'-  
161 CTCGGTACGTTCCGACGC-3'), a probe targeting both *Tb. denitrificans* and *Tb. thioparus*, was  
162 chosen due to satisfactory brightness in FISH. Another probe, TMD131 (5'-  
163 TCCCAGTCTTTGAGGTAC-3'), was developed for hybridising with *Tm. denitrificans*. The  
164 hybridisation conditions of all probes were optimized using pure cultures of the target  
165 microorganisms.

166

### 167 **FISH and CARD-FISH**

168 The biomass samples were analyzed by FISH and CARD-FISH with the aim to compare both  
169 techniques. The higher signal intensity in the hybridisation using CARD-FISH allowed a better cell  
170 detection and quantification by means of image analysis. Few changes in the microbial  
171 composition of the sludge in the two reactors were detected in samples obtained at different times  
172 of operation. The microbiota of both reactors was found to be very similar. Bacteria was the most  
173 abundant Domain (40–60% of the total DAPI-stained cells).  $\gamma$ -*Proteobacteria* (25–45%), in  
174 combination with  $\alpha$ - and  $\beta$ -*Proteobacteria* (less than 10% each), represented the Bacteria almost  
175 in its entirety. It is noteworthy that in the methanogenic granular sludge used as inoculum,  $\alpha$ -  
176 *Proteobacteria* was the predominant group, with only 1% of  $\gamma$ -*Proteobacteria*.

177

178 The specific oligonucleotides developed in this study showed the presence of both autotrophic  
179 denitrifiers. *Tb. denitrificans* was the dominant species, whereas *Tm. Denitrificans* only  
180 represented a minor fraction of the microorganisms (2–3%). Phylogenetically, Thiobacillus is

181 grouped in the  $\beta$ -*Proteobacteria*. Surprisingly, the number of *Tb. denitrificans* cells detected was  
182 very similar to the number of  $\gamma$ -*Proteobacteria* cells. Further tests with pure cultures showed that  
183 *Tb. denitrificans* could hybridise with the  $\gamma$ -*Proteobacteria* probe but not with the  $\beta$ -*Proteobacteria*  
184 probe. This fact can explain the high relative abundance of cells in the  $\gamma$ -*Proteobacteria* class. In  
185 the methanogenic inoculum, *Tb. denitrificans* was detected occasionally but after 110 days of  
186 operation its number increased significantly, from 1% to 35% of the total DAPI-stained cells.  
187 These results suggest that *Tb. denitrificans* was enriched during the operation of the bioreactors  
188 and that this microorganism probably played a dominant role in the autotrophic denitrification.

189

190 Microorganisms in the *Archaea* domain were less abundant and only represented about 10% of  
191 the total DAPI-stained cells, both in the inoculum and in the biomass of the autotrophic reactors.  
192 Only cells from the orders *Methanobacteriales* (1% of the total DAPI-stained cells) and  
193 *Methanosarcinales* (the number significantly decreased from the inoculum to the autotrophic  
194 reactors) were found. The unspecific hybridisation of the *Thiobacillus* cells with the Arc915 probe  
195 (González-Toril, personal comm.) could explain the difference between the numbers of cells  
196 hybridised with the *Archaea* probe and the *Methanobacteriales* plus *Methanosarcinales* probes.

197

## 198 **Cloning**

199 Table 1 summarizes the phylogenetic affiliations of 16S clones corresponding to the inoculum  
200 and to biomass obtained at the end of the operation of R1. The Gram-positive phyla, *Firmicutes*  
201 and *Actinobacteria*, were dominant in the inoculum. Clones from the phylum *Bacteroidete* (clones  
202 E-J-36, E-G-12 and E-Q-101) resemble clone populations found at sites contaminated with  
203 chlorinated solvents. All genera detected include strict or facultative fermentative anaerobes  
204 which is in agreement with the source of the microbial culture, an anaerobic reactor.

205

206 Analysis of the R1 clone library showed that although the diversity of Gram-positive  
207 microorganisms was maintained, species most closely related to the *Proteobacteria* became  
208 dominant. A considerable number of clones (42%) corresponded to former "Pseudomonas",  
209 currently known as *Burkholderiales*, *Xhantomonadales* and *Pseudomonadales*, which are all  
210 microbes well known for their ability to degrade xenobiotic compounds. The only autotrophic

211 denitrifier was *Thiobacillus denitrificans*, which was not detected in the inoculum. Although cloning  
212 is not a quantitative technique, it is interesting to note that the percent *Tb. denitrificans* clones in  
213 the 16S clone library (30%) corresponded closely to the relative abundance of this bacterium as  
214 determined by CARD-FISH (35%). For the *Archaea* domain, 85 clones were obtained which after  
215 restriction digestion with the enzyme Sau3A1 could be grouped into 7 different restriction patterns.  
216 Surprisingly, all the clone sequences were closely related (97–99% similarity) to a single  
217 methanogenic species, *Methanosaeta soehngeni*. In the autotrophic reactors, the methanogens  
218 must be likely resting forms from the inoculum (methanogenic sludge).

219

## 220 **SEM**

221 The structure of representative granules from the inoculum and sludge obtained from both  
222 reactors at the end of the continuous experiment was studied by SEM. Figures 1A and B show  
223 that the methanogenic inoculum had a compact structure with a fluffy surface. With time, the  
224 structure of granules from the denitrifying bioreactors became looser, showing increased voids  
225 (Figures 1C–F). Some granules were hollow (Figures 1F and C detail). Possibly the growth of  
226 microorganisms was restricted to the outer side, giving the granule surface a continuous, dense,  
227 smooth aspect (Figure 1G). Channels perpendicular to the surface were observed (Figures 1C  
228 and D). These effects could cause operation problems in the reactor by increasing the buoyancy  
229 of the sludge, which may result in sludge flotation and/or wash-out problems.

230

## 231 **CONCLUSIONS**

232 FISH oligonucleotide probes for the autotrophic denitrifiers *Thiobacillus denitrificans* and  
233 *Thiomicrospira denitrificans* were designed and their utility was tested. Cloning and CARD-FISH  
234 data showed that bacterial diversity in the two autotrophic denitrifying bioreactors operated with  
235 thiosulfate as electron donating substrate and bicarbonate as carbon source changed during the  
236 first period of operation and later became nearly constant. Bacterial sequences most closely  
237 related to Gram-positives were dominant in the sludge used as inoculum.  $\alpha$ -,  $\beta$ - and  $\gamma$ -  
238 *Proteobacteria* were dominant in the mature autotrophic reactor. The archaeal diversity remained  
239 almost unchanged and was represented by *Methanosaeta soehngeni*.

240



241 *Thiobacillus denitrificans* appeared to be the dominant denitrifier in the two autotrophic  
242 bioreactors examined in this study based on FISH-CARD cell counts and high abundance in the  
243 16S clone library. Some microbial granules became less dense and lost their integrity and  
244 structure during the operation, turning into a shell probably due to the superficial growth of  
245 microorganisms.

246

## 247 REFERENCES

248 Alphenaar, P.A., Groeneveld, N. and Van Aelst, A.C. (1994). Scanning electron microscopical  
249 method of internal structure analysis of anaerobic granular sludge. *Microbiol.*, 25, 129–133.

250

251 Amann, R., Fuchs, B.M. and Behrens, S. (2001). The identification of microorganisms by  
252 fluorescence *in situ* hybridisation. *Curr. Opin. Biotech.*, 12, 231–236.

253

254 Amann, R., Ludwig, W. and Schleifer, K.H. (1995). Phylogenetic identification and *in situ* detection  
255 of individual microbial cells without cultivation. *Microbiol. Rev.*, 59, 143–169.

256

257 Fuchs, B.M., Wallner, G., Beisker, W., Schwippl, I., Ludwig, W. and Amann, R. (1998). Flow  
258 cytometric analysis of the *in situ* accessibility of *Escherichia coli* 16S rRNA fluorescently labelled  
259 oligonucleotide probes. *Appl. Environ. Microbiol.*, 64, 4973–4982.

260

261 Green, M., Schnitzer, M., Tarre, S., Shelef, G., Bilanovic, G. and Soeder, C.J. (1995).  
262 Heterotrophic denitrification using fluidized-bed reactor. *Acta Hydrochim. Hydrobiol.*, 23, 61–65.

263

264 Kuenen, J.G., Robertson, L.A. and Tuovinen, O.H. (1992). The genera *Thiobacillus*,  
265 *Thiomicrospira* and *Thiosphaera*. In *The Prokaryotes*, Vol.2, 2nd edn, Balow, A. (ed.), Springer,  
266 Berlin, Germany, pp. 2638–2657.

267

268 Lane, D.J. (1991). 16S/23S rRNA sequencing. In *Nucleic Acid Techniques in Bacterial*  
269 *Systematic*, Stanckebrandt, E. and Goodfellow, M. (eds), Wiley, New York, USA, pp. 115–175.

270 Muyzer, G., Hottenträger, S., Teske, A. and Wawer, C. (1996). Denaturing gradient gel  
271 electrophoresis of PCR-amplified 16S rDNA – A new molecular approach to analyze the genetic  
272 diversity of mixed microbial communities. In *Molecular Microbial Ecology Manual*, Akkermans,  
273 A.D.L. and van Elsas, J. (eds), Kluwer, Dordrecht, The Netherlands, section 3.4.4, pp. 1–23.  
274

275 Nugroho, R., Takanashi, H., Hirata, M. and Hano, T. (2002). Denitrification of industrial  
276 wastewater with sulfur and limestone packed column. *Wat. Sci. Tech.*, 46(11–12), 99–104.  
277

278 Pernthaler, A., Pernthaler, J. and Amann, R. (2002). Fluorescence *in situ* hybridization and  
279 catalyzed reporter deposition for the identification of marine bacteria. *Appl. Environ. Microbiol.*,  
280 68, 3094–3101.  
281

282 Raskin, L., Strimley, J.M., Rittmann, B.E. and Stahl, D.A. (1994). Group specific 16S RNA  
283 hybridization probes to describe natural communities of methanogens. *Appl. Environ. Microbiol.*,  
284 60(4), 1232–1240.  
285

286 Reyes-Avila, J.S., Razo-Flores, E. and Gomez, J. (2004). Simultaneous biological removal of  
287 nitrogen, carbon and sulfur by denitrification. *Wat. Res.*, 38(14–15), 3313–3321.  
288

289 Robertson, W.D. and Anderson, M.R. (1999). Nitrogen removal from landfill leachate using an  
290 infiltration bed coupled with a denitrification barrier. *Ground Wat. Monit. R.*, 19, 73–80.  
291

292 Schedel, M. and Truper, H.G. (1980). Anaerobic oxidation of thiosulfate and elemental sulfur in  
293 *Thiobacillus denitrificans*. *Arch. Microbiol.*, 124, 205–210.  
294

295 Sierra-Alvarez, R., Guerrero, F., Rowlette, P., Freeman, S. and Field, J.A. (2005). Comparison of  
296 chemo-,hetero- and mixotrophic denitrification in laboratory-scale UASBs. *Wat. Sci. Tech.*, 52(1–  
297 2), 337–342.  
298

299 Stahl, D.A. and Amann, R. (1991). Development and application of nucleic acid probes. In Nucleic  
300 Acid Techniques in Bacterial Systematics, Stackebrandt, E. and Goodfellow, M. (eds), Wiley,  
301 Chichester, UK, pp. 205–248.

302

303 Timmer-Ten Hoor, A. (1981). Cell yield and bioenergetics of *Thiomicrospira denitrificans*  
304 compared to *Thiobacillus denitrificans*. Anton. Leeuw., 47, 231–243.

305

306 Wagner, M., Horn, M. and Daims, H. (2003). Fluorescence *in situ* hybridisation for the  
307 identification and characterisation of prokaryotes. Curr. Opin. Microbiol., 6, 302–309.

308

309 Wang, H.Y. and Qun, J.H. (2003). Combined bioelectrochemical and sulfur autotrophic  
310 denitrification for drinking water treatment. Wat. Res., 37, 3767–3775.

311

312

313

314

315

316

317

318

319

320

321

322

323

324

325

326

327

328

## 329 TABLES

330 Table 1 16S rDNA sequence-based phylogenetic affiliations of the Bacteria clones

| Clone   | Phylum/class/order                                 | Closest species or taxon (NCBI access no.) | Similarity (%) | Abundance (%) |
|---|--|--|----------------|---------------|
| Inoculum (methanogenic sludge                 |  |  |                |               |
| E-N-108                                       | <i>δ-Proteobacteria</i>                            | <i>Desulfovibrio alcoholovorans</i>        | 97             | 2.8           |
| E-E-50  | <i>Firmicutes</i>                                  | Uncultured anaerobic reactor AY426453      | 98             | 5.6           |
|   | <i>Clostridiales</i>                               | <i>Acidaminococcus fermentans</i>          | 93             |               |
| E-P-33  | <i>Firmicutes</i>                                  | Uncultured granular sludge AY426451        | 98             | 2.8           |
| E-B-76  | <i>Actinobacteria</i>                              | Uncultured anaerobic reactor AB195904      | 99             | 39.1          |
|   | <i>Actinomycetales</i>                             | <i>Propionibacterium</i>                   | 92             |               |
| E-I-30  | <i>Actinobacteria</i>                              | <i>Cellulomonas fermentans</i>             | 99             | 8.4           |
|   | <i>Actinomycetales</i>                             |  |                |               |
| E-M-93  | <i>Actinobacteria</i>                              | Uncultured granular sludge AY426443        | 98             | 2.8           |
|   | <i>Actinomycetales</i>                             | Uncultured <i>Rubrobacteridae</i> AY395379 |                |               |
| E-D-59  | <i>Chloroflexi</i>                                 | Uncultured anaer. consortium AJ009475      | 95             | 5.6           |
|   |  | Uncultured AY304372                        | 94             |               |
| E-J-36  | <i>Bacteroidete</i>                                | Uncultured from aquifer AF050541           | 100            | 22.3          |
| E-G-12  | <i>Sphingobacteriales</i>                          | <i>Cytophaga</i> AB015525                  | 94-97          | 8.4           |
| Reactor biofilm (alter 186 days of operation) |  |  |                |               |
| R1C-M-20                                      | <i>α-Proteobacteria</i>                            | <i>Ochrobactrum</i>                        | 99             | 9             |
|   | <i>Rhizobiales</i>                                 | ( <i>O. tritici</i> )                      |                |               |
| R1C-A-5                                       | <i>β-Proteobacteria</i>                            | <i>Thiobacillus denitrificans</i>          | 94-98          | 30            |
| R1C-C-27                                      | <i>Hydrogenophilales</i>                           |  |                |               |
| R1C-E-91                                      |  |  |                |               |
| R1C-G-86                                      | <i>β-Proteobacteria,</i><br><i>Burkholderiales</i> | <i>Comamonas</i>                           | 97             | 6             |
| R1C-L-1                                       | <i>γ-Proteobacteria</i>                            | <i>Rhodanobacter</i>                       | 95-98          | 30            |
| R1C-D-73                                      | <i>Xhantomonadales</i>                             | ( <i>R. lindaniclasticus</i> )             |                |               |
| R1C-B-96                                      |  |  |                |               |
| R1C-J-90                                      | <i>γ-Proteobacteria,</i><br><i>Pseudomonadales</i> | <i>Pseudomonas</i>                         | 95             | 6             |
| R1C-H-6                                       | <i>Firmicutes,</i><br><i>Clostridiales</i>         | <i>Clostridium</i>                         | 95             | 6             |
| R1C-K-79                                      | <i>Firmicutes, Bacillales</i>                      | <i>Exiguobacterium (E. aestuari)</i>       | 98             | 3             |
| R1C-P-24                                      | <i>Firmicutes</i>                                  | Uncultured anaer. digester UEU81756        | 98             | 3             |
|   | <i>Lactobacillales</i>                             | <i>Alkalibacterium</i>                     | 91             |               |
| R1C-O-10                                      | <i>Chloroflexi</i>                                 | Uncultured AY921924                        | 97             | 3             |
|   | <i>Anaerolineales</i>                              | <i>Caldilinea</i>                          | 93             |               |
| R1C-N-108                                     |  | Uncultured intestinal microflora AY916338  | 97             | 3             |
|   |  | Uncultured bioreactor AY050603             | 97             |               |

331

332

333

334

335

336

337

338

339

340

341

342

343 **FIGURE LEGENDS**

344 **Figure 1** Scanning electron microphotographs of granules. A: Methanogenic granular sludge  
345 used as inoculum; B: Detail of the surface of the inoculum; C: Granule from the autotrophic  
346 denitrifying reactor R1 after 186 d of operation. Detail in the bottom-left corner shows a hollow  
347 granule; D: Detail of the radial structures formed in the granule showed in C; E: Granule from the  
348 autotrophic denitrifying reactor R2 at the end of operation; F: Outer shell of these granules; G and  
349 H: Details of the surface and inside of granules from R2. Morphologies resembling *Thiobacillus*  
350 can be observed. Note the densely packed biofilm on the surface and the heterogeneity inside  
351 the granule. Detail in E shows a sludge granule from R2 established on an active carbon particle

352

353

354

355

356

357

358

359

360

361

362

363

364

365

366

367

368

369

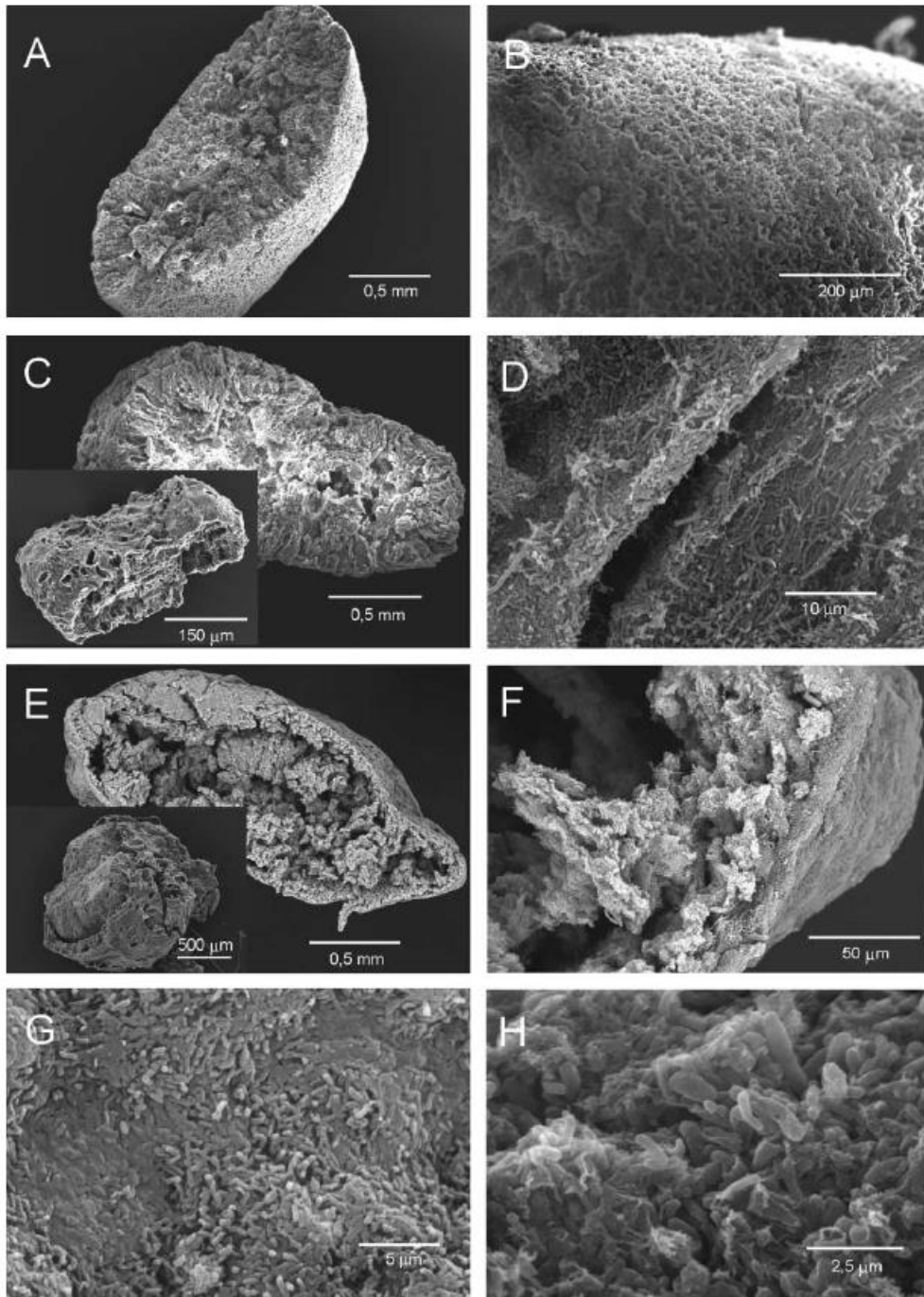
370

371

372

373 FIGURES

374 Figure 1



375